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**TECHNICAL
REPORT**

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**Dose conversion factors for major
nuclides within high level waste**

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Studsvik Nuclear

November 1990

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DOSE CONVERSION FACTORS FOR MAJOR NUCLIDES WITHIN
HIGH LEVEL WASTE

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This report concerns a study which was conducted for SKB. The conclusions and viewpoints presented in the report are those of the author(s) and do not necessarily coincide with those of the client.

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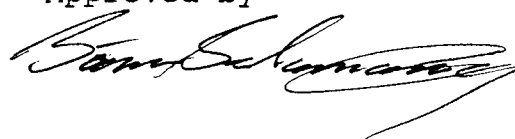
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Approved by

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ABSTRACT

Individual doses to critical groups from a continuous unit release of nuclides to a standard biosphere were calculated. The selection of nuclides for this study was based on experience of their importance from a radiological point of view. The standard biosphere consisted of a well and a lake with adjacent farming-land. It was assumed that 1 % of the activity reached the well directly, while the remaining fraction was directly diluted into the lake water. Ten exposure pathways for activity from the well and the lake water to man were considered. The ecosystem was assumed to be similar to present conditions in Sweden. This was also the case concerning diet and living habits. In addition the doses from naturally occurring nuclides in the uranium decay chains were calculated, based on natural levels in water and soil. The calculations were carried out with the BIOPATH and PRISM codes. The latter code was used to obtain the uncertainty in the results due to the uncertainty in the input parameter values.

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INTRODUCTION

The individual doses from unit releases of long-lived radioactive nuclides contained within a deep geological repository were calculated and are reported in /Bergström et al, 1990/. The doses were calculated for a so called standard biosphere which constituted of a well and lake with adjacent farming-land. Current habits and metabolic conditions of human beings were assumed. Adults and five year old children were considered, respectively. The calculations were carried out with the emphasis to study the importance of uncertainties in the input parameter values. Discussion of the results from these uncertainty analyses are going to be published /Bergström et al, 1991/. One major contribution to the uncertainty in the results was the dilution volume for the nuclides in the groundwater. This dilution was studied by varying reservoir volumes as well as varying the fraction of activity reaching the water in the well. For some radionuclides additional calculations were carried out for a fixed fraction of the nuclides reaching the well /Bergström et al, 1991/.

In this summary report conversion factors between unit releases and doses to adults are presented for most nuclides appearing in considerable amounts in high level waste. In addition some nuclides are handled which also belong to the natural decay chain of uranium. For these nuclides doses are also calculated based upon natural occurring concentrations of them in water and soil.

In the release case the calculations are carried out with a basic assumption that a fraction of 1 % of each nuclide reaches the well directly. All other assumptions are likewise those reported in /Bergström et al, 1990/. A brief description of the model used, the exposure pathways considered and the used values of the input parameters are given in the report, for a more detailed description, see /Bergström et al, 1990/.

The code BIOPATH /Bergström et al, 1982/ was used for calculations of the turnover of radionuclides in the biosphere and the PRISM-system /Gardner et al, 1983/ for obtaining the ranges of the uncertainty due to the uncertainty and variability in input parameter values.

NUCLIDES

Conversion factors were calculated for all radiologically important nuclides contained in high level waste. In addition factors were calculated for those nuclides belonging to the decay-chain of natural uranium but not appearing substantially in high level waste. Contributions from daughter nuclides were considered if they contributed significantly to the total dose. This is the case if the daughter nuclides are so long-lived that a considerable amount can be generated during the time studied. The nuclides treated, half-times and dose factors are shown in Table 2-1. The dose factors are the sum of weighted committed organ dose equivalents according to ICRP-standards.

Table 2-1 Nuclides assessed, half-lives and dose factors (Sv/Bq).

Nuclide	Half-life (years)	Inhalation	Ingestion
C-14	5.7E3	5.6E-10 1)	5.6E-10 1)
Se-79	6.4E4	2.4E-9 2)	2.3E-9 2)
Sr-90	2.9E1	6.0E-8 1)	3.5E-8 1)
Zr-93	1.5E6	8.6E-8 2)	4.2E-10 2)
Nb-93m	1.4E1	7.7E-9 2)	1.4E-10 2)
Tc-99	2.1E5	2.0E-9 2)	3.4E-10 2)
Sn-126	1.0E5	2.3E-8 2)	4.7E-9 2)
I-129	1.6E7	4.0E-8 1)	6.4E-8 1)
Cs-135	2.3E6	1.2E-9 2)	1.9E-9 2)
Cs-137	3.0E1	8.6E-9 1)	1.3E-8 1)
Pb-210	2.2E1*	3.4E-6 2)	1.4E-6 2)
Po-210	1.4E2*	2.2E-6 2)	5.0E-7 4)
Ra-223	1.1E1*	2.0E-6 2)	1.5E-7 2)
Ra-225	1.5E1*	2.0E-6 2)	3.1E-7 2)
Ra-226	1.6E3	2.1E-6 2)	3.1E-7 2)
Ac-227	2.2E1	1.8E-3 2)	3.8E-6 2)
Th-229	7.3E3	5.7E-4 2)	9.4E-7 2)
Th-230	7.7E4	8.6E-5 2)	1.6E-7 4)
Pa-231	3.2E4	3.4E-4 2)	2.2E-5 3)
U-233	1.6E5	3.6E-5 2)	3.1E-7**
U-234	2.5E4	3.6E-5 2)	3.0E-7**
U-235	7.0E8	3.3E-5 2)	2.8E-7 4)
U-236	2.3E7	3.4E-5 2)	2.9E-7**
U-238	4.5E9	3.2E-5 2)	2.7E-7 4)
Np-237	2.1E6	5.5E-5 1)	4.5E-7 1)
Pu-239	2.4E4	1.2E-4 1)	9.7E-7 1)
Pu-240	6.5E3	1.2E-4***	9.7E-7***
Pu-241	1.4E1	2.3E-6 1)	1.9E-8 1)
Pu-242	3.8E5	1.1E-4***	8.8E-7***
Am-241	4.3E2	1.1E-4 1)	8.9E-7 1)

* Given in days.

** Based upon values for U-235 and U-238 given in Johansson, 1984.

*** Based upon values for Pu-239 and Pu-241 given in ICRP56.

References

- 1) ICRP56
- 2) ICRP30
- 3) Johansson, 1982
- 4) Johansson, 1984

METHODOLOGY

The calculations were performed assuming a continuous leakage of 1 Bq per year of each nuclide in soluble form to the biosphere during a period of 500 years. Of the leakage 1 % was assumed to enter the well water directly while the remaining fraction leaked directly to the lake water. No delay or reduction of activity by accumulation in the interphase geosphere and biosphere was considered. With the objective to simulate the exposure, a seven compartment model of the studied biosphere was designed, see Figure 1 where also masses are given and flows of activity are described by arrows. The BIOPATH-code was used for solving the differential equations and calculating the doses. For the nuclides belonging to the natural occurring uranium decay chain doses were calculated for the same exposure pathways based upon the concentrations given in Table 3-1. These are based upon a brief literature survey, see Tables A.1 and A.2, and represent average background values.

The uncertainty in the results due to the uncertainty in input parameter values were examined with the PRISM-system. Some general data of interest are given in Appendix A, Table A.3.

The uncertainty analyses were carried out for each nuclide. All parameter values with the exception of dose conversion factors and volumes were varied.

Table 3-1 Nuclides assessed in the natural uranium decay chains. Mean concentration and ranges as logtriangularly distributed extracted from /UNSCEAR 1988; Bowen, 1979; Yengar, 1990; Sundblad et al, 1985; Landström et al, 1986; Hallstadius et al, 1984; Eriksson et al, 1981 Kulich et al, 1988/.

Nuclide	Soil (Bq/kg)		Water (mBq/l)	
	Mean	Range	Mean	Range
U-238	50	10 - 100	10	1 - 100
U-234	50	10 - 100	10	1 - 100
Th-230	50	10 - 10	10	1 - 100
Ra-226	80	10 - 200	10	1 - 100
Po-210	100	10 - 1000	10	1 - 100
Pb-210	100	10 - 1000	10	1 - 100
U-235	2	0.1 - 10	0.5	0.05 - 5
Pa-231	2	0.1 - 10	0.5	0.05 - 5
Ac-227	2	0.1 - 10	0.5	0.05 - 5
Ra-223	2	0.1 - 10	0.5	0.05 - 5

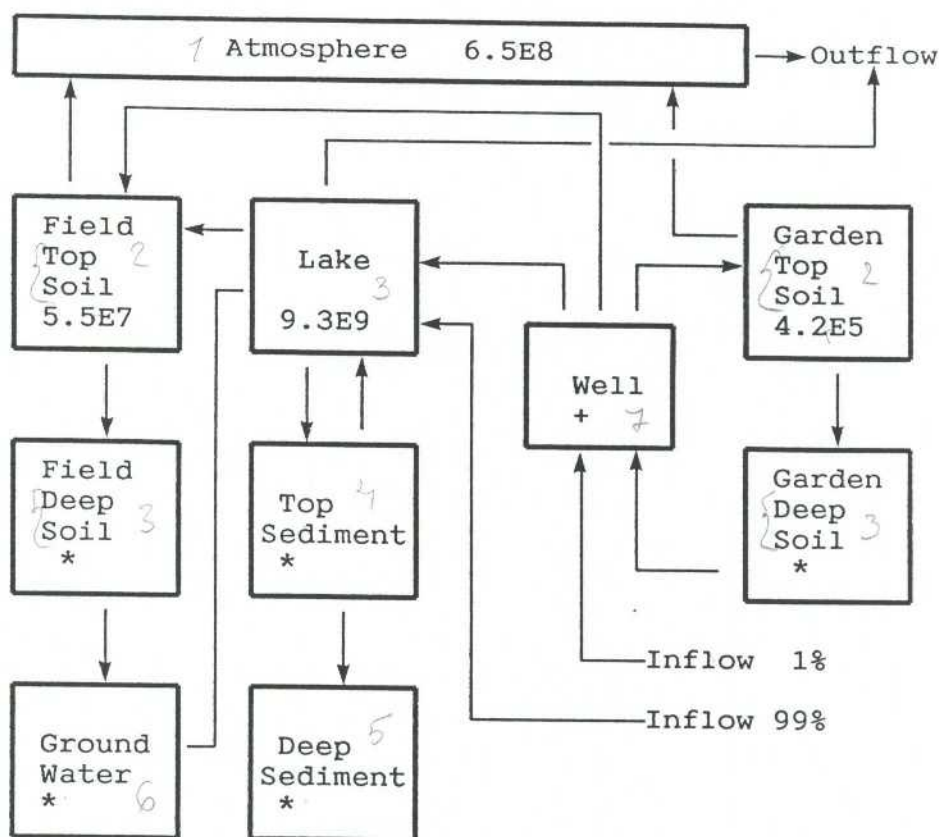


Figure 3-1 The structure of the compartment model with masses (kg).

* Not of importance for the results.

+ The volume is varied within the calculations.

For the naturally occurring radionuclides the garden plot and field had the same background levels. This was also assumed concerning the levels in ground and lake water.

The pathways for ingestion of nuclides are via different types of food and drinking water. The intake of soil is also included. This latter pathway is valid via e.g. consuming unwashed vegetables. This intake is adopted to 10 g/y mainly from the garden plot.

Earlier calculations of the doses from these long-lived nuclides showed that the internal exposure dominates the exposure for the nuclides considered.

The only external exposure considered is from ground. This represents staying on the fields and the garden plots. The dose factors used are taken from /Svensson, 1979/ with assumption of homogenous distribution of the activity in soil.

The following pathways were considered:

- Water
- Milk
- Meat
- Vegetables
- Root vegetables (potatoes)
- Cereals
- Fish
- Soil
- External exposure from ground

The pathways considered and the compartments they emanate from are shown in Figure 3-2.

All biological uptake parameters such as root-uptake factors, bioaccumulation factors to fish and steady state factors giving the concentration in milk and meat from continuous intake are shown in Appendix A, Tables A.4 to A.6. The values used are based upon a literature survey carried out and reported in /Bergström et al, 1990/. The Tables A.4 to A.6 also show references for those nuclides not studied in that report.

Consumption data used are given in Appendix A. Table A.7.

In Tables A.4 to A.6 references are also given for the values used for those nuclides not handled in the earlier report.

In addition, some more nuclides are treated which whose references are given the tables.

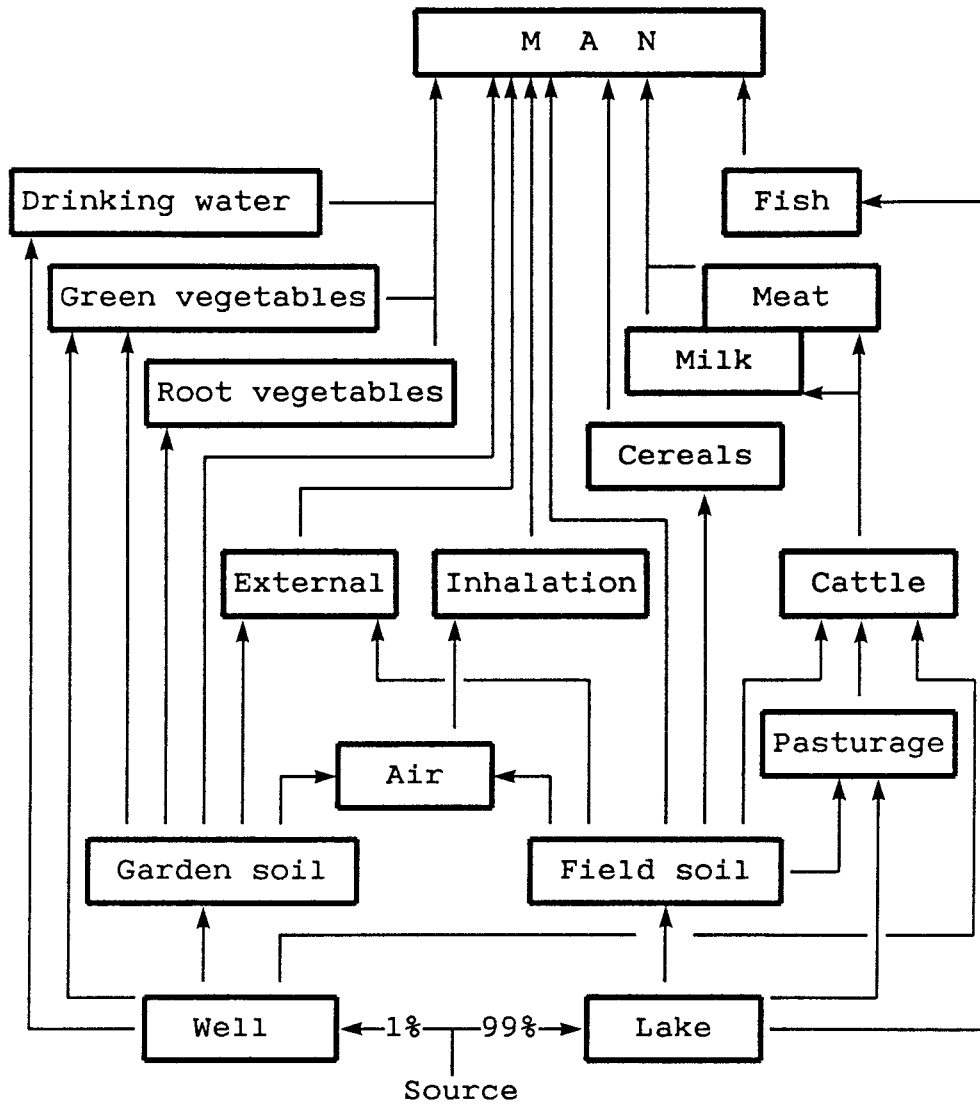


Figure 3-2 Exposure pathways for the critical group.

RESULTS

Results, as arithmetic mean values of the total dose are presented in Table 4-1. All these results do not consider any contributions from daughter nuclides. Such contributions were studied for Zr-93, Th-229 and Th-230, because of experience from earlier calculations /Bergström, 1989/. The daughter products studied are Nb-93m, Ra-225 and Ra-226 respectively. These contributions were only notable for Zr-93 and Th-229. Including them the conversion factors would increase with 7 and 36 percent, respectively.

The percentual contributions to the total dose from dominant exposure pathways are given in Table 4-2. In the table the percentual contribution to the total uncertainty from the respective exposure pathway is given within brackets.

Dominant pathway is drinking water for most nuclides, see Table 4-2. The fraction of activity reaching the well is constant in these calculations. That implies that the estimated ranges of uncertainty are quite small, see Table 4-1.

In Table 4-3 the annual doses from the naturally occurring nuclides are given as well as the doses given in UNSCEAR 88 for respective nuclide. The percentual contribution to the dose via different pathways are shown in Table 4-4. There are differences between the "release" case and the natural case concerning the dominant exposure pathways. This is due to the steady-state conditions the natural case reflects. The relation between the concentration of the nuclides in well water compared to soil and lake water is different in the two cases. The concentration in the well water is relatively higher in the "release" case which results in drinking of this as dominant exposure pathway.

Table 4-1 Individual doses from unit releases to adults arithmetic mean and ranges corresponding to 2.5 and 97.5 percentiles (Sv/Bq).

Nuclide	Arithm mean	Ranges
C-14	1.3E-14	(0.6 - 2.3)E-14
Se-79	6.3E-14	(1.6 - 37)E-14
Sr-90	2.2E-13	(1.4 - 4.1)E-13
Zr-93	2.5E-15	(1.6 - 4.3)E-15
Tc-99	1.7E-15	(1.2 - 2.5)E-15
Sn-126	3.7E-14	(2.2 - 7.6)E-14
I-129	5.4E-13	(3.1 - 13)E-13
Cs-135	3.9E-14	(1.2 - 17)E-14
Cs-137	2.4E-13	(0.7 - 9.9)E-13
Pb-210	6.7E-12	(4.6 - 9.9)E-12
Po-210	8.1E-13	(6.1 - 11)E-13
Ra-223	3.5E-14	(2.3 - 6.6)E-14
Ra-225	9.2E-14	(6.1 - 17)E-14
Ra-226	1.8E-12	(1.1 - 3.4)E-12
Ac-227	1.7E-11	(1.1 - 2.4)E-11
Th-229	4.4E-12	(3.1 - 6.6)E-12
Th-230	7.5E-13	(4.9 - 12)E-13
Pa-231	1.0E-10	(0.6 - 1.5)E-10
U-233	1.6E-12	(1.1 - 2.5)E-12
U-234	1.6E-12	(1.0 - 2.4)E-12
U-235	1.5E-12	(1.0 - 2.3)E-12
U-236	1.5E-12	(1.1 - 2.3)E-12
U-238	1.4E-12	(1.0 - 2.2)E-12
Np-237	2.2E-12	(1.4 - 3.3)E-12
Pu-239	4.3E-12	(3.1 - 6.1)E-12
Pu-240	4.3E-12	(2.9 - 6.2)E-12
Pu-241	7.9E-14	(5.7 - 11)E-14
Pu-242	3.9E-12	(2.6 - 5.6)E-12
Am-241	4.2E-12	(2.8 - 6.8)E-12

Table 4-2 Percentual contribution from dominant exposure pathways to the total dose and, within brackets, to the uncertainty.

Nuclide	Drinking water	Milk	Meat	Vegetables	Potatoes	Fish
C-14	18 (1)	8 (6)	3 (-)	3 (4)	- (-)	68 (91)
Se-79	15 (-)	1 (-)	- (-)	19 (18)	49 (80)	15 (1)
Sr-90	60 (11)	2 (-)	- (-)	29 (72)	7 (15)	1 (-)
Zr-93	63 (14)	- (-)	22 (75)	13 (10)	- (-)	1 (-)
Tc-99	81 (57)	- (-)	1 (-)	15 (41)	1 (1)	1 (1)
Sn-126	50 (8)	5 (1)	1 (-)	16 (35)	11 (39)	17 (16)
I-129	46 (4)	21 (20)	1 (-)	17 (58)	4 (12)	9 (3)
Cs-135	19 (-)	5 (-)	5 (-)	6 (-)	5 (-)	60 (99)
Cs-137	20 (-)	5 (-)	6 (-)	4 (-)	1 (-)	64 (100)
Pb-210	77 (47)	1 (-)	- (-)	14 (43)	- (-)	9 (10)
Po-210	81 (58)	- (-)	1 (2)	16 (40)	- (-)	2 (-)
Ra-223	69 (8)	15 (83)	- (-)	12 (8)	- (-)	4 (1)
Ra-225	69 (8)	15 (82)	- (-)	12 (8)	- (-)	3 (1)
Ra-226	67 (4)	16 (84)	- (-)	14 (6)	3 (5)	1 (-)
Ac-227	82 (59)	- (-)	- (-)	15 (36)	- (-)	3 (5)
Th-229	80 (49)	- (-)	- (-)	16 (46)	3 (5)	- (-)
Th-230	80 (45)	- (-)	- (-)	16 (48)	3 (7)	- (-)
Pa-231	82 (51)	- (-)	1 (-)	15 (46)	1 (3)	- (-)
U-233	76 (51)	1 (-)	8 (16)	14 (33)	- (-)	1 (1)
U-234	76 (48)	1 (-)	8 (20)	14 (31)	- (-)	1 (-)
U-235	76 (48)	1 (-)	8 (20)	14 (31)	- (-)	1 (-)
U-236	76 (49)	1 (-)	8 (21)	14 (30)	- (-)	1 (-)
U-238	76 (51)	1 (-)	8 (16)	14 (32)	- (-)	1 (1)
Np-237	81 (54)	- (-)	1 (-)	15 (41)	2 (4)	1 (1)
Pu-239	85 (62)	- (-)	- (-)	15 (38)	- (-)	- (-)
Pu-240	84 (56)	- (-)	- (-)	15 (43)	- (-)	- (-)
Pu-241	85 (62)	- (-)	- (-)	15 (38)	- (-)	- (-)
Pu-242	84 (56)	- (-)	- (-)	15 (43)	- (-)	- (-)
Am-241	79 (39)	- (-)	- (-)	17 (55)	- (1)	3 (5)

Table 4-3 Individual annual doses to adults from naturally occurring uranium and daughter nuclides in soil and water, arithmetic mean and ranges corresponding to 2.5 and 97.5 percentiles (Sv/y) and corresponding doses reported in UNSCEAR for respective nuclide.

Nuclide	Arithmetic	Ranges	UNSCEAR 88
U-238	7.3E-5	(0.5 - 31)E-5	5E-6
U-234	8.4E-5	(1 - 39)E-5	
Th-230	8.4E-6	(1 - 35)E-6	7E-6
Ra-226	1.3E-4	(0.1 - 7.2)E-4	7E-6
Po-210	2.8E-5	(0.4 - 9.6)E-5	
Pb-210	5.3E-4	(0.5 - 23)E-4	1.2E-4
U-235	3.3E-6	(0.3 - 17)E-6	
Pa-231	4.3E-5	(0.5 - 15)E-5	
Ac-227	2.6E-5	(0.2 - 13)E-5	
Ra-223	1.7E-6	(0.2 - 7)E-6	

Table 4-4 Percentual contribution from dominant exposure pathways to the total dose and within brackets, to the uncertainty for naturally occurring uranium nuclide and decay products.

Nuclide	Water	Milk	Meat	Vege- tables	Potatoe	Cereals	Fish
U-238	3 (-)	1 (-)	7 (5)	2 (1)	2 (2)	72 (88)	13 (3)
U-234	3 (-)	- (-)	7 (3)	2 (1)	2 (1)	72 (93)	13 (1)
Th-230	17 (15)	- (-)	1 (1)	11 (8)	11 (5)	3 (2)	5 (72)
Ra-226	2 (-)	40 (74)	1 (-)	12 (6)	17 (9)	19 (9)	9 (1)
Po-210	17 (16)	2 (1)	12 (7)	7 (7)	6 (5)	7 (6)	44 (56)
Pb-210	2 (-)	1 (1)	- (-)	4 (4)	11 (9)	58 (77)	22 (7)
U-235	4 (1)	1 (-)	7 (6)	2 (1)	2 (1)	69 (85)	15 (4)
Pa-231	23 (15)	- (-)	2 (1)	6 (6)	7 (7)	32 (30)	27 (41)
Ac-227	7 (2)	- (-)	- (-)	10 (1)	1 (-)	- (-)	81 (97)
Ra-23	4 (1)	26 (33)	1 (1)	13 (14)	19 (26)	19 (21)	17 (5)

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Appendix A.1

Table A-1 Natural radionuclides in soil (Bq/kg).

Nuclide	Sundblad	Bowen	Eriksson		UNSCEAR	Hallstadius	Landström
			Mean	St dev			
U-238	21-220	24 (8-110)	70	57	25	3-29	6-434
U-234		26 (9-120)	77	61	25		
Th-230	100	100 (3700-16000)*	62	107	25		
Ra-226	39-120	(30 (7-180)	82	96		15-34	40-295
Po-210		8-220					
Pb-210		75-6300*				20 (12-1000)	

* From one abnormal soil containing 750-3000 Bq U-238/kg.

Table A-2 Natural radionuclides in water (mBq/l).

Nuclide	Bowen	UNSCEAR	Hallstadius	Sundblad	Kulich
U-238	4.8	25 (0.1 - 50)	0.3 - 47	5 - 36	
U-234	5.2		0.4 - 80		
Th-230	-				
Ra-226	4 - 400	22 (7 - 1800)	0.7 - 20	1 - 29	2 - 2455**
Po-210	0.5 - 2.6				
Pb-210	3 - 8		2 - 24		

** Ground-water from private wells.

Appendix A.2

Table A-3

Some general input parameter values.

Parameter	B.E.	Type of distr*	Min	Max
Annual demand of water for man (m ³)	50	T	30	70
Annual demand of water for the live-stock (m ³)	230	T	200	260
Irrigation of the garden plot (l/m ² year)	150	T	30	300
Irrigation of agricultural land (l/m ² year)	150	T	30	300
Daily consumption of foodstuff for cattle (kg d w)	14	T	12	16
Daily consumption of soil when grazing (kg)	0.3	T	0.1	0.5
Residence time of water in the lake (year)	0.75	T	0.6	0.9
Turnover time of groundwater (year)	5	T	1	10
Area of garden plot (m ²)	1000	C		
Area of agricultural land (m ²)	130000	C		

* T = Triangular distribution.
C = Constant.

Appendix A.3

Table A-4

Root uptake factors for several types of nutrients (Bq/kg f w nutrient per Bq/kg d w soil).

Element	Distribution*	Pasturage (dw/dw)	Ranges or geom st dev	Cereals	Ranges or geom st dev	Vegetables	Ranges or geom st dev	Root vegetables	Ranges or geom st dev	
Se	LT	6.5	5E-1 - 7E1	1.5	1E-1 - 8E1	6.5	5E-1 - 6E1	1.3E1	1 - 1E2	
Sr	LT	3.2	1 - 7	1.4	0.5 - 3	5E-1	0.2 - 1.0	1.5-1	8E-2 - 0.3	(Bergström, 1990)
Zr**	LT	8E-4	1E-4 - 1E-2	2E-4	1E-5 - 1E-3	2E-4	1E-5 - 1E-3	2E-4	1E-5 - 1E-3	(Bergström, 1988)
Tc	LT	1.0	1E-1 - 1E1	9E-1	1E-1 - 1E1	2E-1	1E-2 - 1.0	1E-1	1E-2 - 1.0	
Sn	LT	1E-1	1E-2 - 1.0	4E-1	1E-2 - 1.0	6E-2	1E-2 - 1.0	5E-2	1E-2 - 1.0	
I	LN	6E-1	4.0	1E-1	4.0	3E-2	4.0	1E-2	4.0	
Cs	LN	5E-2	2.4	1E-2	1.8	2E-2	1.9	2E-2	1.9	
Pb	LT	2E-2	1E-3 - 1E-1	2E-2	1E-3 - 1E-1	2E-3	1E-4 - 1E-2	4E-3	1E-3 - 1E-2	
Po**	LT	4E-3	4E-4 - 4E-2	2E-4	2E-5 - 2E-3	2E-4	2E-5 - 2E-3	2E-4	2E-5 - 2E-3	(IAEA, 1982)
Ra	LN	5E-2	2.5	1E-2	2.5	1E-2	2.5	1E-2	2.5	
Ac	LT	5E-4	3E-5 - 7E-3	4E-4	1E-5 - 1E-3	4E-3	2E-4 - 8E-2	5E-5	2E-5 - 1E-2	
Th	LT	1E-2	1E-3 - 1E-1	7E-4	1E-4 - 1E-3	2E-3	1E-4 - 1E-2	2E-3	1E-4 - 1E-2	
Pa	LT	3E-3	3E-4 - 3E-2	3E-3	3E-4 - 3E-2	3E-4	3E-5 - 3E-3	6E-4	6E-5 - 6E-3	
U	LT	1E-2	1E-3 - 1E-1	4E-2	4E-3 - 4E-1	1E-3	1E-4 - 1E-2	1E-3	1E-4 - 1E-2	
Np	LT	1E-1	1E-2 - 1	1E-2	1E-3 - 1E-1	3E-3	5E-4 - 2E-1	3E-3	5E-4 - 2E-1	
Pu	LT	1E-3	7E-5 - 1E-2	1E-4	1E-6 - 1E-2	2E-5	1E-7 - 4E-4	5E-5	2E-7 - 1E-3	
Am***	LT	5E-4	3E-5 - 7E-3	4E-4	1E-5 - 1E-3	4E-3	2E-4 - 8E-2	5E-5	2E-5 - 1E-2	

* LT = Logtriangular distribution.
LN = Lognormal distribution.

** Assumed ranges.

*** Same as for Ac-227, see Bergström 1990.

Appendix A.4

Table A-5

Bioaccumulation factors to fish (Bq/kg f w muscle per Bq/l).

Element	Best estimate	Type of distr*	Geom st dev	Ranges	References
C	4600	T		3000-6000	
Se	2000	T		500 - 5000	
Sr	5	T		1 - 20	(Bergström, 1990)
Zr**	60	T		10 - 100	(Bergström, 1988)
Tc	15	T		1 - 50	
Sn	3000	T		1000 - 6000	
I	200	T		10 - 500	
Cs	5000	LN	3.8		
Pb	100	T		50 - 200	
Po**	50	T		10 - 100	(IAEA, 1982)
Ra	50	T		10 - 100	
Ac	100	T		10 - 1000	
Th	30	T		1 - 100	
Pa	10	T		1 - 100	
U	50	T		10 - 100	
Np	50	T		1 - 100	
Pu	5	T		1 - 50	
Am***	100	T		10 - 1000	

* T = Triangular distribution.
LN = Lognormal distribution.

** Estimated ranges.

*** Same as for Ac-227, see Bergström, 1990.

Appendix A.5

Table A-6

Distribution factors for transfer to milk and meat, logtriangularly distributed.

Element	Milk (day/l)	Ranges or geom st dev	Meat (day/kg)	Ranges or geom st dev	References
C	1E-2	1E-3 - 1E-1	3E-2	1E-3 - 1E-1	
Se	3E-3	1E-3 - 1E-2	9E-4	1E-4 - 1E-2	
Sr	8E-4	4E-4 - 3E-3	6E-4	7E-5 - 1E-3	(Bergström, 1990)
Zr**	5E-6	5E-7 - 5E-5	3E-2	3E-3 - 3E-1	(Bergström, 1988)
Tc	1E-4	1E-5 - 1E-3	2E-3	1E-4 - 1E-2	
Sn	3E-3	1E-3 - 1E-2	1E-3	1E-4 - 1E-2	
I*	1E-2	1.6	2E-3	2.1	
Cs*	8E-3	1.6	3E-2	2.1	
Pb	3E-4	2E-5 - 2E-3	4E-4	4E-5 - 4E-3	
Po**	1E-4	1E-5 - 1E-3	3E-3	1E-4 - 1E-2	(IAEA, 1982)
Ra*	3E-3	3.9	7E-4	1.2	
Ac	3E-7	3E-8 - 3E-6	1E-5	1E-6 - 1E-4	
Th	5E-6	1E-7 - 1E-4	7E-4	1E-4 - 1E-3	
Pa	5E-5	1E-6 - 1E-4	3E-3	2E-6 - 5E-3	
U	2E-4	2E-5 - 2E-3	1E-2	1E-3 - 1E-1	
Np	5E-6	1E-6 - 1E-4	3E-3	2E-4 - 5E-3	
Pu	1E-7	2E-8 - 3E-7	2E-6	1E-7 - 2E-5	
Am***	3E-7	3E-8 - 3E-6	1E-5	1E-6 - 1E-4	

* Lognormal distribution.

** Estimated ranges.

*** Same as for Ac-227, see Bergström, 1990.

Appendix A.6

Table A-7

Consumption and habit data, triangular distribution.

	Best estimate	Min	Max
<u>Individuals</u>			
Inhalation (m ³ /y)	8000	7000	9000
Drinking water (l/y)	600	450	750
Milk (l/y)	200	20	400
Meat (kg/y)	55	5	100
Green vegetables (kg/y)	40	5	100
Cereals (kg/y)	80	5	150
Root-fruits (kg/y)	70	5	150
Fish (kg/y)	30	5	100
Soil (kg/y)	0.01	0.001	0.1
Exposure time, garden plot (h/y)	200	100	400
Exposure time, fields (h/y)	200	100	400
Being outdoors (h/y)	1000	500	1500
Filtration factor, building	0.3	0.1	0.5

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²SKB AB

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University of Gothenburg, Department of General and Marine Microbiology, Gothenburg
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Lars Werme¹, Patrik Sellin¹, Roy Forsyth²
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²Battelle, Chicago
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²INTERRA/ECL, Leicestershire, UK
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A. B. MacKenzi³, L. Moreno⁴, I. Neretnieks⁴,
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